



Inter-Basin Water Transfers: Balancing Water Scarcity Solutions with Environmental and Socio-Economic Impacts from Nepalese Perspective

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Abstract

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Inter-basin water transfer (IBWT) involves moving water across drainage divides to address water scarcity, support agriculture, industry, recreation, and generate hydropower. While practiced since ancient times, modern IBWTs gained momentum in the 19th century and have since expanded globally. These projects range from small-scale transfers to large water transfer megaprojects (WTMPs). IBWTs can help mitigate water shortages, enhance food and energy security, and support ecosystem restoration, but also pose risks such as habitat degradation, biodiversity loss, and socio-economic disruptions. Though Nepal has practiced IBWT since the 17th century, the complex environmental and socio-economic consequences observed globally have yet to be fully assessed in the country. Furthermore, a number of IBWTs in the country are in pipeline. Therefore, this review is an attempt to shed light on different aspects of IBWTs including a brief historical perspective; the scale; and to evaluate the environmental and; socio-economic impacts globally and describe the scenario and associated implications in Nepalese context. This study selected articles based on IBWTs, their relevance to environmental impacts, socio-economic effects, and the role of IBWTs. Sources include peer-reviewed journals, government reports, and case studies, focusing on global examples and specific challenges in Nepal. The major findings of this study is that IBWTs are gaining popularity and are becoming important in water-food-energy nexus despite their environmental and socio-economic implications. However, if such projects are undertaken with comprehensive environmental assessments, sustainable water management practices, with inclusive policy frameworks, countries can leverage IBWT projects to meet their growing demands for water, energy, and food, while safeguarding ecological integrity and community welfare.

Keywords: Environmental impacts, inter-basin water transfer, socio-economy

Introduction

Freshwater ecosystems cover less than 1% of Earth's surface, yet it is incredibly diverse encompassing more than 400 large-scale ecoregions harbouring at least 10% of the earth's species (Grooten & Almond, 2018). Inland waters and freshwater biodiversity constitute a valuable natural resource, in economic, cultural, aesthetic, scientific and educational terms (Arya, 2021). The global distribution of freshwater shows tremendous spatio-temporal variation (Rodell et al., 2018; Wetzel, 2001) attributed to seasons, locations, total precipitation events (Qin et al., 2019; Zhang et al., 2014) along with the magnitude and frequency of extreme climatic events (Kalyan et al., 2021; Yu & Ma, 2022). The present world is facing a lot of challenges to ensure the access of sufficient water resources because of the increasing dependency of human societies on water (Larsen et al., 2016); and non-uniform distribution of freshwater (Somlyódy & Varis, 2006) and climate change (Heino et al., 2009). This is further exacerbated by humans' developmental activities (Cosgrove & Loucks, 2015; Pittock et al., 2009). Humans have explored and developed several ways to minimize the issues of water scarcity such as recycling wastewater, damming rivers, groundwater extraction, cloud seeding, seawater desalination, virtual water trade, inter-basin water transfer, adoption of rain water harvesting technologies, and restoration of wetlands (Hutchinson et al., 2010; Opare, 2012; Zhuang, 2016). Out these different practices, the concept

between river basins and implementation of IBWT peaked in the 1980s. IBWT is defined as “the *purposeful arrangement of natural hydrologic patterns via engineering works (dams, reservoirs, tunnels and pumping stations) to move water across drainage divides to satisfy human and other needs*” (Micklin, 1984). Therefore, IBWTs have been recognized as engineering solutions to store, redistribute and treat water resources and involves transfer of excess water over extended distances (Rollason et al., 2021) from geographically separated water-surplus basins to water deficit basins (Golubev & Biswas, 1978; Snaddon et al., 1999) to cater to water needs for irrigation, power generation, industrial development and recreation (Grant et al., 2012; Pittock et al., 2009; Rollason et al., 2021).

Although IBWTs prove to be beneficial, these projects come with price as they can have a number of environmental and socio-economic impacts (Liu et al., 2023). Environmental impacts include change upstream change in water quality; biotic assemblages; hydromorphology whereas the socio-economic impacts include water conflicts, community displacement associated with loss of livelihood (Annys et al., 2019). Despite their environmental and socio-economic implications, IBWTs are on the rise. In this context, this review attempts to shed lights on the brief history of IBWTs, their scales; associated environmental and; socio-economic implications of IBWTs globally with focus in Nepalese context as the latter has a number of IBWTs in pipeline. For this review, a range of peer-reviewed articles; national and international reports were cited relevant to IBWTs.

IBWTs: Past and present

Although IBWTs as modern engineering interventions peaked only in the 1980s (Rollason et al., 2021), different modes of water transfer were already in practice in ancient civilizations in Egypt, Jericho in Jordan, China during 3100 -2100 BC (Liang & Greene, 2019; Snaddon et al., 1999). In ancient Mesopotamia, it was practiced during the Bronze Age as early as 4000-11000 BC where several canals were connected from the Euphrates River for regions of Sumer and Akkad (Mays, 2010; Tamburrino, 2010). In those days, different water transfer practices included transfer through canals, aqueducts, underground cisterns, rainwater harvesting etc. (Table 1)

Table 1: Sources of water for cities of the early civilizations (4000–1100 B.C.)

Sources of water	Cities (palaces)
Short canal connected to permanent river	Uruk, Ur, Babylon (all cities in the Tigris and Euphrates valleys)
Canals and reservoirs storing flood water of nonpermanent river, rainfall	Jawa, Khirbet el Umbashi
Rainwater harvesting (gutters and cisterns)	Agia Triadha, Chamaizi, Mari, Knossos, Myrtos-Pyrgos, Phaistos, Zakross
Wells	Ugarit (Syria), Palaikastro, Knossos, Zakros, Kommos, Mohenjo Daro (Indus Valley)
Aqueducts from source at altitude	Knossos*, Mallia*, Tylissos, Pylos, Thebes, Dur Untawsh (Elam)
Underground cisterns w/ steps	Mycenae, Athens, Tyrins, Zakros, Tylissos
Springs	Knossos, Tylissos, Syme

Source: Ancient Water Technologies 2010.

Australia, the USA and India are considered as the 19th century pioneers of inter-basin water transfer. In Australia, the Goulburn Valley Irrigation Scheme implemented in 1886 is considered as one of the earliest irrigation schemes in the country aimed at diverting water from the Goulburn River to irrigate the fertile Goulburn Valley (Oldham & Moody, 1913). Likewise, in India, the Ganges Canal, completed in 1854, diverted water from the Ganges River to irrigate the fertile Doab region between the Ganges and Yamuna rivers (Lata, 2019). The Erie Canal in the USA completed on October 26, 1825, connected the Great Lakes with the Hudson River, facilitating transportation and commerce (Morton & Olson, 2019). Since then, a number of IBWTs have been developed and implemented in several other countries as well including Israel, Canada (Gleick, 2000; Shiklomanov & Rodda, 2004). In 2005, approximately 14% of the total water withdrawal from rivers in the world involved IBWT and the development of IBWTs; and water transfer in near future is expected to rise by

different scales being operated in about 40 countries and regions across the world, with the total annual water transfer amounting to around 500 billion m³ (Su & Chen, 2021).

Scales of IBWT

Scales of IBWT varies in terms of distance, regions, amount of water being transferred; accordingly, IBWTs maybe long-distance water transfer, inter-regional water transfer, large-scale water transfer, inter-catchment water transfer, inter-basin water transfer and intra-basin water transfer (Fang et al., 2015; Golubev & Biswas, 1978; Purvis & Dinar, 2020; Shumilova et al., 2018; Verma et al., 2009). In recent times, a newer concept on IBWT scale – water transfer megaprojects (WTMPs) - have been developed based on construction costs (>USD1 billion); distance of transfer (>190 km) or volume of water (>0.23 km³ per year) (Shumilova et al., 2018). There are about 34 existing and 76 future (planned, proposed or under construction) WTMPs globally focusing on agricultural, domestic supply, hydropower development, mining, ecosystem restoration and transformational routes (Shumilova et al., 2018). Some of the notable WTMPs include the California State Water Project (SWP) which is one of the largest state-built water and power development and conveyance systems in the United States. Initiated in 1960 and managed by the California Department of Water Resources with an estimated cost of 9 billion US\$, it involved the construction of 1128 km long canal, transferring about of 3.33 km³a⁻¹ of water. It provides water to over 27 million people and irrigates approximately 750,000 acres of farmland. It supports both urban and agricultural areas with generating hydroelectric power, contributing to the state's energy supply and helping to offset the project's operational costs (Sabet & Coe, 1986).

Likewise, the South-to-North Water Transfer Project (SNWTP) in China is one of the world's largest and most ambitious water diversion projects. It aims to connect the southern Yangtze River and northern Yellow River with a total of 2,700 miles tunnels and canals via three distinct routes through western, central and eastern China. The SNWTP has already significantly improved water availability in northern China, supporting urban and industrial growth and improving living standards in water-scarce regions (Berkoff, 2003; Miao et al., 2018). The National River Interlinking Project in India is another ambitious WTMP. It was first proposed in the 1970s by K.L. Rao but the project officially started to take shape only in 2002. It aims to divert a staggering 174 Billion m³ of water through a canal network of 14900 km with an estimated cost US\$120 Billion. It aims to connect 37 rivers across the nation through a network of nearly 3000 storage dams to build a gigantic South Asian Water Grid. This project is expected to irrigate around 35 million hectares of land, raising the ultimate irrigation potential from 140 million hectares to 175 million hectares and generation of 34000 megawatts of hydropower, apart from the incidental benefits of flood control, navigation, water supply, fisheries, salinity and pollution control (Alagh et al., 2006; Joshi, 2013). Considering the current global population growth and scarcity of land; but the growing needs of food, water and electricity for growing populations, IBWTs and WTMPs are gaining popularity and are becoming important in water-food-energy nexus.

Impacts of IBWT

IBWTs have been considered beneficial because of their multi-purpose uses and benefits (Laassilia et al., 2021; Sun et al., 2023). However, the current state of knowledge indicates that large dams, along with their positive impacts, inter-basin transfers and water withdrawal have a number of negative impacts on environment (Snaddon et al., 1999; Zhuang, 2016) as well as on economy and communities (Flyvbjerg, 2014; Gupta & van der Zaag, 2008). Any IBWT system can result in complex physical, chemical, hydrological and biological implications for both the donor and receiving basins (Davies et al., 1992). The Aswan High Dam (AHD) of the Nile in Egypt undoubtedly offers prime example of how river damming and diversion have complicating impacts (Biswas & Tortajada, 2011; Kashef, 1981; Zeid, 1989). The nature and the extent of the impacts vary widely depending on the type and characteristics of water transfer, biophysical and socio-economic factors. Positive impacts include adding new basins for water deficient areas (Purvis & Dinar, 2020; Shao et al., 2003), facilitating water cycle (Yano et al., 2018), improving meteorological conditions in the recipient basins (Khadem et al., 2021; Murgatroyd & Hall, 2020), mitigating ecological water shortage (Duan et al., 2022), repairing the damaged ecological system and preserving the endangered wild fauna and flora (Dadaser-Celik et al., 2009; Wang et al., 2014), generating hydroelectric power (Erskine et al., 1999), controlling flood (Khadem et al., 2021), irrigation (Wang et al., 2021), transport routes (Liang & Greene, 2019) and water recreation and tourism (Akron et al., 2017). However, the magnitude of recent water transfer projects in response to fast increased demand has overlooked the severe ecological, environmental, economic and social risks associated with water transfer (Daga et al., 2020; Li et al., 2017). Concerns about the environmental, societal and economic

consequences of inter-basin water transfers has been raised in recent periods (Laassilia et al., 2021; Pittock et al., 2009; Wilson et al., 2017; Zhang et al., 2015). A large number of studies have examined the negative impacts of inter-basin water transfers on ecosystem impacts and socio-economic disruptions such as forest depletion, soil and water contamination, waterborne diseases, livelihood, human displacement and migration (Bui et al., 2020; Das, 2006; Gallardo & Aldridge, 2018; Gleick, 1993). Following sections briefly summarize the environmental and socio-economic impacts of IBWTs.

Environmental impacts

Environmental impacts of IBWT are mostly attributed to change in natural flow (Zhuang, 2016) and include change in different physico-chemical and biological parameters of both donor and recipient basins and water bodies. Global reviews on the negative impacts of IBWTs have revealed implications on terrestrial dynamics; biodiversity and water quality (Ghassemi & White, 2007; Snaddon et al., 1998; Snaddon et al., 1999; Zhuang, 2016).

Impacts on physico-chemical parameters

Changes in natural flow affects a range of water physico-chemical parameters such as water temperature, salinity, turbidity, erosion, sedimentation, waterlogging, mineral and nutrient concentrations, hydrology, oxygenation, inorganic substrate composition, sediment dynamics, land use changes in both donor and recipient basins (de Lucena Barbosa et al., 2021; Gallardo & Aldridge, 2018; Marak et al., 2020; Tian et al., 2019). The change in the flow conditions of rivers and lakes in the receiving basin although can lead to increased water levels but it can also cause potential flooding in areas adjacent to the water bodies (de Lucena Barbosa et al., 2021). Furthermore, changes in flow can also affect the pattern and magnitude of sediment transport and deposition (Hamidifar, 2024). Likewise, changes in natural flow affect riparian ecosystem health as it diminishes the water bodies' ability to assimilate pollutants and thus cause pollution, eutrophication, salinization and acidification (Zhuang, 2016). The transfer of water between basins can also introduce new geochemical elements and compounds into the receiving basin. This can affect the mineral composition of the water and sediments, potentially leading to changes in the aquatic chemistry (Jiao et al., 2021). For instance, water transfer from the Yellow River to the Fen River in China has resulted in increase in the concentrations of Na^+ and Cl^+ ions along with increased conductivity values in the Fen River (Yuan et al., 2020). Apart from these, IBWTs result changes in water levels; and renewal rates decline in downstream main channels (Pittock et al., 2009); disrupt river connectivity and; flood plains and channels connectivity (Bunn & Arthington, 2002; Grant et al., 2012). Changes in water transparency, nutrient and sediment loads, channel morphology and granulometry are some of the long-term physico-chemical effects of dams on environments downstream (Granzotti et al., 2018; Kamidis et al., 2021; Szatten et al., 2021; Yang et al., 2021), potentially leading to long-term nutrient loading (He et al., 2020; Stockner et al., 2000).

Impact on aquatic organisms

The changes in the physico-chemical parameters associated with IBWTs in turn affect the biological parameters and almost all groups of biotas are affected (Rehman et al., 2015; Sharma et al., 2016). Sediment deposition can result in excessive algal growth thereby compromising water quality resulting in eutrophication; affect the food web; alter habitats for organisms (Glibert & Burford 2017; Li et al., 2023). The changes in water quality and hydrology can have cascading effects on the aquatic ecosystems and the associated biota (Li et al., 2023).

Fish assemblages are one of the most affected communities due to IBWTs. A large number of studies have revealed change in fish assemblages both in the donor and recipient basins (Gallardo & Aldridge, 2018; Schmidt et al., 2020). IBWTs affect fish assemblages through isolation, alteration and degradation of habitats, blockade of migratory routes, change in nutrient concentrations and food webs, flow and temperature (Ghassemi & White, 2007; Snaddon et al., 1998). Habitat alteration and degradation are resulted because of the disruption of the river continuum (Doretto et al., 2020; Ward & Stanford, 1995) followed by drowning of channel and erosion of riparian habitats which often act as fish spawning pockets (Ghassemi & White, 2007). In addition, introduction of invasive species during IBWTs and spread of diseases also affect the fish assemblages. There are several reports of loss of native species and establishment of invasive/exotic species which are likely to be established in altered habitats (Dudgeon et al., 2006; Gallardo & Aldridge, 2018). Isolation of fish assemblages and nonulations can increase competition among the resident species for food and breeding sites (Andrades et

al., 2021; Belmaker et al., 2005). It may also lead to decrease in genetic diversity and therefore puts species at greater risk from disease, altered predation pressure, behavioral changes and increased vulnerability to environmental changes (Coleman et al., 2018; Kano et al., 2016; Lymbery et al., 2020; Ruzich et al., 2019). Furthermore, long-term isolation could lead to interspecific hybridization and thus have serious consequences on fish biota genetic diversity particularly the native species (Allendorf et al., 2001). Studies suggest that reduction in the nutrient concentrations in reservoirs due to damming affect food webs resulting in the change in the structure of primary producer communities, detritivores as well as the consumers (Maavara et al., 2020; Macura et al., 2019; Wang et al., 2022). If fish movement occurs, it can lead to invasion or altered regional connectivity patterns, spread of non-native species which could induce biotic and genetic homogenization or synchronization (Li et al., 2022; Schmidt et al., 2020; Shao et al., 2019). In contrast, habitat alterations may favour introduce invasive species in recipient basins. For instance, five new fish species (*Labeobarbus aeneus*, *Clarias gariepinus*, *Labeo capensis*, *Austroglanis sclateri* and *Labeo umbratus*) have been transferred to the Great Fish River in the Eastern Cape from the Orange Orange/Vaal River (Zhang et al., 2015). Likewise, fish species like *Gobio gobio*, has been introduced into the Segura River from the donor Tajo River in Spain; *Catostomus fumeiventris*, was transferred to the Los Angeles Basin from northern donor rivers (Snaddon et al., 1999). The introduced species can become invasive and eliminate native fauna through predation, competition, and higher reproductive success (Mayfield et al., 2021). Furthermore, the invasive species can modify the behaviour (such as habitat usage, diel activity) of the native species. Thus, invasive species can have negative impacts on the native species attributed to disruption of food webs, loss of biodiversity, hybridization and spread disease (Bernery et al., 2022; Ellender & Weyl, 2014; Olden et al., 2022).

Impacts on macroinvertebrates include loss of headwater species (Clarke et al., 2008; Guerold et al., 2000), hindering of macroinvertebrate passage along a stream/river stretch (Guareschi et al., 2014), decreased macro-benthic diversity (Rolls et al., 2012), change in food webs (Murphy et al., 2019; Panikkar et al., 2021; van der Zee et al., 2016) and eventually altering the community composition (Ko et al., 2020). For instance, dominant Chironomidae, Hydropsychidae and Simuliidae taxa were replaced by *Simulium chutteri* in Great Fish River from Orange River, South Africa (O'keeffe & De Moor, 1988). Change in relative abundance of different functional feeding groups of macroinvertebrates have also been reported by several authors with increased abundance of scrapers and collector filterers (Brittain & Saltveit, 1989; Vallania & Corigliano, 2007). Thus, it is evident that IBWTs affect macro-benthic communities by changing the latter's assemblages.

Socio-economic impacts

Apart from the environmental impacts, IBWTs also have upstream as well as downstream socio-economic impacts (Liu et al., 2023; Snaddon et al., 1998). The socio-economic impacts vary from individual impacts to entire community and society attributed to construction work, influx of people and increased fringe urbanization (Mutanga et al., 2013). Individual impacts include loss of one's property and livelihood whereas impacts on communities include displacement of people and their homes, disturbance and loss of local livelihoods, loss of productive farmland, loss of cultural heritage sites and monuments, health hazards etc. (Das, 2006; Pittock et al., 2009; Zhuang, 2016).

Upstream communities relying on agriculture sector on donor basins often face negative consequences on their agricultural economy due to reduced quality and quantity of water (Gichuki & McCornick, 2008). Conflicts may arise mostly due to disputes in water sharing though other factors such as increased pollution has also been reported as a causal factor (Guardiola-Avila et al., 2018; Madani et al., 2011). Such conflicts may be global, regional or local depending on the scale, extent and impacts of IBWTs (Hernández-Mora et al., 2014; Yevjevich, 2001). For instance, construction of a dam on Ethiopian Nile resulted in conflicts between the neighbouring countries of Egypt, Sudan and other upstream nations (Madani et al., 2011); construction of the Farakka Barrage on the Ganges has resulted conflicts between Bangladesh and India as it compromised water availability and water demand for Bangladesh (Islam, 2012). The South-North Water Diversion Project (SNWDP) is one of the most recent transboundary water disputes over water sharing of the Brahmaputra River between the neighbouring countries of India and China (Ho et al., 2019).

IBWTs in Nepalese Scenario: Past, Present and Future

In the context of Nepal, the history of river diversion dates back to 17th century during the Malla regimes, when

1990; Parajuli & Sharma, 2003). A number of such canals still exist in different parts of the country such as Tikabhairav, Bageswori, Budhikanta Canal etc (Shrestha & Dahal, 2020). The first large water transfer project was constructed in 1923 AD and completed in 1928 AD which drew water from the Triyuga River in Udaipur District (Pradhan, 1989). However, the idea of large water transfer projects dates back only to the 1970s when the Government of Nepal commissioned a study to explore the Babai Irrigation Project to irrigate fertile land on the western plains of Nepal during the non-monsoon period. It involved the construction of highway weir cum bridge over the Babai River at Parewa Odar in western Nepal. A 5.5 km canal was constructed and it started feeding the traditional canals -Budhi *kulo*, Majro *Kulo*, Raj *Kulo* and Dhadhwar *Kulo* - and the water irrigated about 4000 ha land on the western plains during the non-monsoon period (GoN/BIP, 2001). Similarly, the Kulekhani Storage Hydropower project is the only storage-type hydropower project of Nepal which is the first project that transfers water from Kulekhani river of Bagmati basin to East Rapti river of Gandaki Basin (Pradhan et al., 2012).

The Government of Nepal (GoN) guided by the National Planning Commission has initiated more than 20 ambitious infrastructural developmental projects as National Pride Projects to enhance the quality of life in terms of social, economic, cultural and environmental aspects (GoN/NPC, 2022). These projects are strategically important for the development of different sectors viz. hydroelectricity, irrigation, transportation, tourism, cultures & religion, etc. Infrastructural development to improve socio-economic status of the country is a national agenda in Fifth Year Plan (GoN/NPC, 2020). With rich freshwater resources (WECS, 2011), Nepal has huge potential for hydropower generation, the and to expand irrigation (ADB, 2018; GoN/DWRI, 2019). Irrigation is given a third priority by the Water Resources Act (1992) after the use of water for drinking and domestic purposes. This clearly signifies the importance of irrigation to boost agriculture and achieve food security. Thus, water resource-based infrastructural development is being considered as an important component of food-water-energy nexus in the country. The Bheri-Babai Diversion Multi-purpose Project (BBDMP) is one such national pride project and construction began only in 2015 where water from the Bheri River is being diverted to the Babai River via a 12.3 km transfer with design flow of 40.0 m³/s (GoN/BBDMP, 2018). The Sunkoshi Marin Diversion Multipurpose Project (SMDMP) is another IBWT project and the construction began in 2022 where water from the Sunkoshi River to the Marin River, a tributary of the Bagmati River transferred via a 13.3 km long tunnel which generate 31.07 MW of electricity (GoN/SMDMP, 2022)(GoN/SMDMP, 2022). Likewise, the Melamchi Water Supply Project (MWSP) was initiated by the Government of Nepal to divert a water volume of 1,70,200 m³ /day through a 26 km underground tunnel from the Melamchi River (Koshi Basin) to Kathmandu Valley (Bagmati Basin) to ease the chronic water shortage situation within the Kathmandu Valley (Bhattarai et al., 2005). The project has been delivering water in the Valley through a temporary structure for certain months in a year. A number of IBWT projects in the country are in pipeline focusing on irrigation and hydropower generation (GoN/DWRI, 2019) (Table 2). However, the likely environmental and socio-economic impacts of these projects are yet to be observed and assessed. Migratory routes of fish species like *Tor putitora*, *Tor tor*, *Bagarius bagarius*, *Chupisoma gaura* and *Anguilla bengalensis* from many rivers have been reported to be affected by damming in the country (ADB, 2018). A recent baseline study on fish assemblages of the Bheri and the Babai at the diversion and release sites respectively failed to capture migratory species like *Anguilla bengalensis* implies that migratory routes of the species may well have already affected by the Babai Dam Weir cum Bridge at Parewa Odar at the Babai River constructed in 1993 (Khatri et al., 2024)

Table 2: Some IBWTs in Nepalese context

Name of Project	Feature	Donor/Recipient Rivers/basins
Karnali Diversion Project	19 km length tunnel (59 m ³ /s), 80MW, for irrigation of about 46,000 ha	Karnali River to Mohana River
Madi-Dang Diversion Project	25 km length tunnel (24 m ³ /s), 61MW, for irrigation of about 17,000 ha	Madi River to Dang valley
Naumure Dam: Rapti-Kapilvastu Diversion Project	23 km length tunnel, 343MW, for irrigation of about 40,849 ha	West Rapti River to Kapilvastu
Kaligandaki-Tinau Diversion Project	25 km length tunnel (66 m ³ /s), 244MW, for irrigation of about 62,000 ha	Kaligandaki River to Rupandehi District,
Kaligandaki Nawalparasi Diversion Project	6 km length tunnel (17 m ³ /s), 4mw, for irrigation of about 11,500 ha)	Kaligandaki River-Nawalparasi-East District
Trishuli Shaktikhor Diversion Project	18 km tunnel (51m ³ /s) (No hydropower production), for irrigation of about 13,000 ha.	Trishuli River to Chitwan District
Sunkoshi Diversion Project	a). 14 km length tunnel (77 m ³ /s), 33MW, for irrigation of about 55,000 ha,	Sunkoshi River to the Marin Rivers
	b). 17 km length tunnel (72 m ³ /s), 44MW, for irrigation of about 129,000 ha	Sunkoshi River to the Kamala Rivers
Tamor Morang Diversion Project	a). 31 km length tunnel, 90MW, for irrigation of about 45,000 ha in Morang District	Tamor River to the Chisang River
Kankai Multipurpose Project	Dam at Kankaimai River of 85-meter height, 80MW, for irrigation of about 40,000 ha in Jhapa District	Kankaimai to Jhapa District
Chatara Barrage Project	Construction of Barrage at Koshi River (No hydropower production), for irrigation of about 66,000 ha on Saptari District.	Koshi River at Chatara to Saptari District
Seti-Pandul Diversion Project	42 km length tunnel, 280 MW), for irrigation of about 300,000 ha.	Karnali Basin

Conclusion

Inter-basin water transfers (IBWT) serve as crucial solutions for addressing water scarcity, electricity generation, supporting agricultural and industrial demands, enhancing energy security and socio-economic upliftment. However, the review highlights that these projects often come with significant trade-offs, including habitat disruption, biodiversity loss, and socio-economic displacement. While IBWTs contribute to regional development by redistributing water resources, they can also lead to habitat alteration, biodiversity loss, and socio-economic disruptions, particularly affecting upstream and downstream communities. In Nepal, the implementation of IBWTs dates back to the 17th century and categorization of recent water resource-based infrastructural developments as national pride projects by the government suggests that roles IBWTs would play in socio-economy of the country is significant. However, the global evidence of IBWT's complex environmental and socio-economic consequences necessitates careful consideration before implementation in the Nepalese context. This review underscores the need for comprehensive environmental assessments, sustainable water management practices, and inclusive policy frameworks to mitigate the risks associated with IBWT projects. By balancing the potential benefits with the associated challenges, countries can leverage IBWT projects to meet their growing demands for water, energy, and food, while safeguarding ecological integrity and community welfare.

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